METHYL ISOAMYL KETONE CAS # 110-12-3 ORAL RISK ASSESSMENT DOCUMENT



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AUTHORS, PEER REVIEWERS, AND ACKNOWLEDGEMENTS

Author:

Toxicology Services Department 1.800.NSF.MARK NSF International 789 Dixboro Road Ann Arbor, MI 48105

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Internal NSF Peer Reviewers:

Lori Bestervelt, Ph.D. Gwendolyn Ball, Ph.D. Clif McLellan, M.S. Maryann Sanders, M.S.

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Michael Dourson, Ph.D., DABT (Vice Chairman, NSF Health Advisory Board) Director TERA (Toxicology Excellence for Risk Assessment)

David Blakey, D.Phil. Director, Environmental Health Science Safe Environments Programme Health Canada

Randy Deskin, Ph.D., DABT Director, Toxicology and Product Regulatory Compliance Cytec Industries, Inc.

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Robert Hinderer, Ph.D. Director of Health, Toxicology, and Product Safety Noveon, Inc.

Jennifer Orme-Zavaleta, Ph.D. Associate Director for Science USEPA/NHEERL/WED

Adi Pour, Ph.D. Director, Douglas County Health Department Omaha, Nebraska

Calvin Willhite, Ph.D. Department of Toxic Substances Control State of California This is a preview of "Methyl isoamyl keton...". Click here to purchase the full version from the ANSI store.

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EXECUTIVE SUMMARY

Methyl Isoamyl Ketone – Oral Risk Assessment CAS # 110-12-3							
PARA	LEVEL	UNITS	CALCULATED:				
NOAEL (no-observed-adverse	25	mg/kg-day	From a 13-week rat inhalation study.				
Oral RfD (oral reference dose)	0.008	mg/kg-day	From a 13-week rat inhalation study.				
TAC (total allowable conce	0.06	mg/L	For a 70 kg adult drinking 2 L/day with a 20% Relative Source Contribution for drinking water.				
SPAC (single product allow	0.006	mg/L	For a 70 kg adult drinking 2 L/day.				
STEL (short term exposure	level)	0.8	mg/L	For a 10 kg child drinking 1 L/day.			
KEY STUDIES	Katz, G.V., E.R. Renner Jr., and C.J. Terhaar. 1986. Subchronic inhalation toxicity of methyl isoamyl ketone in rats. Fund. Appl. Toxicol. 6:498-505. Katz, G.V. 1983. Two week and 90-day inhalation studies of methyl isoamyl ketone in rats. Health and Environment Laboratories, Eastman Kodak Company.						
EFFECT							
UNCERTAINTY FACTORS	 Factors applied in calculating the oral RfD: 3x for interspecies extrapolation 10x for intraspecies extrapolation 10x for subchronic to chronic extrapolation 1x for extrapolation from LOAEL to NOAEL 10x for database deficiencies 						
TOXICITY SUMMARY CONCLUSIONS	Toxicology evaluations of methyl isoamyl ketone include acute, subacute, subchronic, and genotoxicity studies. Developmental and neurotoxicity studies are available for structurally related ketones. The gavage studies located for methyl isoamyl ketone contained deficiencies considered to impact the risk assessment. A 13-week study evaluated only one dose level in male rats only, and a three-week study evaluated only male rats and did not include complete clinical chemistry or histological evaluation. The critical study was a 13-week rat inhalation study, which identified a NOAEL of 212 ppm (human equivalent oral dose of 25 mg/kg-day) based on increased mean absolute and relative liver weight and hepatocyte hypertrophy in both sexes. The changes in liver weight and liver pathology were also observed in the subchronic gavage study. Histopathological changes noted in the kidneys of males were associated with alpha-2µ-globulin nephropathy, thus were not relevant to human health. Based on a rat hepatic peroxisome proliferation study, the liver changes are not likely attributable to hepatic peroxisome proliferation, thus were considered relevant for this risk assessment. A human equivalent oral RfD of 0.01 mg/kg-day was derived from the 13-week rat NOAEL, using appropriate dose conversion factors and an inhalation absorption factor of 50%, based on absorption data for methyl ethyl ketone, since oral or inhalation kinetic, dynamic, or metabolic data for methyl isoamyl ketone were unavailable. Methyl isoamyl ketone was not mutagenic in <i>Salmonella typhimurium</i> over a range of doses in the presence and absence of metabolic activation, and was negative in a 3T3 cell transformation assay. The evidence of chromosomal aberrations observed at high concentrations was discounted since these concentrations were greater than 10 mM and cytotoxicity was observed.						
Conclusions	ketone exposure require data suggests that meth studies were identified f isoamyl ketone <i>cannot</i> action levels established	ad to induce ayl isoamyl for methyl is be determine for methyl i	hepatotoxicity ketone is not oamyl ketone. ed. Taking into isoamyl ketone	in rats. Although the weight of evidence of genotoxicity genotoxic, no chronic animal or human epidemiological Thus the cancer risk to humans from exposure to methyl account the uncertainty factors used, the drinking water are considered to be protective of public health.			

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1.0 INTRODUCTION

This document has been prepared to allow toxicological evaluation of the unregulated contaminant **methyl isoamyl ketone** in drinking water, as an extractant from one or more drinking water system components evaluated under NSF/ANSI 61 (2002), or as a contaminant in a drinking water treatment chemical evaluated under NSF/ANSI 60 (2002). Both non-cancer and cancer endpoints have been considered, and risk assessment methodology developed by the U.S. Environmental Protection Agency (U.S. EPA) has been used.

Non-cancer endpoints are evaluated using the reference dose (RfD) approach (Barnes and Dourson, 1988; Dourson, 1994; U.S. EPA, 1993a), which assumes that there is a threshold for these endpoints that will not be exceeded if appropriate uncertainty factors (Dourson et al., 1996) are applied to the highest dose showing no significant effects. This highest dose is derived from human exposure data when available, but more often is derived from studies in laboratory animals. Either the no-observed-adverse-effect level (NOAEL) taken directly from the dose-response data, or the calculated lower 95% confidence limit on the dose resulting in an estimated 10% increase in response (the LED₁₀ or BMDL from benchmark dose programs) can be used (U.S. EPA, 2001). The lowest-observed-adverse-effect level (LOAEL) can also be used, with an additional uncertainty factor, although the benchmark dose approach is preferred in this case. The RfD is expressed in mg/kg-day. It is defined by the U.S. EPA as "an estimate (with uncertainty spanning perhaps an order of magnitude) of a daily exposure to the human population (including sensitive subgroups) that is likely to be without an appreciable risk of deleterious effects during a lifetime" (Barnes and Dourson, 1988; U.S. EPA, 1993a; U.S. EPA, 1999a).

NSF uses the RfD to derive three product evaluation criteria for non-cancer endpoints. The total allowable concentration (TAC), generally used to evaluate the results of extraction testing normalized to static at-the-tap conditions, is defined as the RfD multiplied by the 70 kg weight of an average adult assumed to drink two liters of water per day. A relative source contribution (RSC), to ensure that the RfD is not exceeded when food and other non-water sources of exposure to the chemical are considered, is also applied in calculating the TAC. The relative source contribution should be data derived, if possible. Alternately, a 20% default contribution for water can be used (U.S. EPA, 1991a). The TAC calculation is then as follows:

$TAC (mg/L) = [RfD (mg/kg-day) \times 70 kg] - [total contribution of other sources (mg/day)]$ 2L/day

or

$$TAC (mg/L) = \frac{RfD (mg/kg-day) \times 70 kg}{2L/day} \times 0.2 (RSC)$$

The single product allowable concentration (SPAC), used for water treatment chemicals and for water contact materials normalized to flowing at-the-tap conditions, is the TAC divided by the estimated total number of sources of the substance in the drinking water treatment and distribution system. In the absence of source data, a default multiple source factor of 10 is used.